Post-Fire Erosional Processes in the Pacific Northwest and Rocky Mountain Regions

Steven M. Wondzell^{a,*} and John G. King^b

^aPacific Northwest Research Station, Olympia Forestry Sciences Lab, 3625 93rd Ave., S.W., Olympia WA 98512 ^bRocky Mountain Research Station, Boise Forestry Sciences Lab, 316 E. Myrtle, Boise ID 83702

Abstract:

The objective of this paper is to provide a general overview of the influence of wildland fires on the erosional processes common to the forested landscapes of the western United States. Wildfire can accelerate erosion rates because vegetation is an important factor controlling erosion. There can be great local and regional differences, however, in the relative importance of different erosional processes because of differences in prevailing climate, geology and topography; because of differences in the degree to which vegetation regulates erosional processes; and because of differences in the types of fire regimes that disrupt vegetative cover. Surface erosion, caused by overland flow, is a dominant response to wildfire in the Interior Northwest and Northern Rocky Mountains (Interior region). A comparison of measured post-fire infiltration rates and longterm records of precipitation intensity suggest that surface runoff from infiltration-excess overland flow should also occur in the Coastal and Cascade Mountains of the Pacific Northwest after fires, but this has not been documented in the literature. Debris slides and debris flows occur more frequently after wildfire in the Interior region and in the Coastal and Cascade Mountains of the Pacific Northwest (Pacific Northwest region). Debris flows can be initiated from either surface runoff or from soil-saturation-caused debris slides. In the Pacific Northwest region, debris flows are typically initiated as debris slides, caused by soil saturation and loss of soil cohesion as roots decay following fire. In the Interior region, both overland-flowcaused and debris-slide-caused debris flows occur after wildfire. Surface erosion, debris slides, and debris flows all occur during intense storms. Thus, their probability of occurrence depends upon the probability of intense storms occurring during a window of increased susceptibility to surface erosion and mass wasting following intense wildfire.

Keywords: Erosion, Fire, Sediment, Over-land flow, Debris Slides, Debris Flows, Ravel

1. Introduction

The objective of this paper is to provide a general overview of the influence of wildland fires on the erosional processes common to the forested landscapes of the western United States. We build upon several recent reviews of the effects of fire on hydrology, geomorphic processes and aquatic ecosystems (Swanson 1981, Beschta 1990, McNabb and Swanson 1990, Gresswell 1999, Wondzell 2001). We examine the physical mechanisms driving erosion, sediment transport and deposition and examine the effects of fire on these erosional processes. We illustrate typical erosional processes, and the influence of fire on those processes, by

comparing and contrasting the Coastal and Cascade Mountains of the Pacific Northwest (Pacific Northwest region) with the forested regions of the Interior Northwest and the Northern Rocky Mountains (Interior region). Our review and synthesis of the published literature is intended to introduce geomorphologic concepts to those in other fields, and to stand as an introduction to other papers in this issue. In depth examination of specific erosional processes is available from the literature cited in this paper, and from other papers in this volume (see Benda, *this issue*; Meyer and Pierece, *this issue*; and Miller et al., *this issue*).

2. Erosion, Transport And Depositional Processes

Erosional processes occur along a continuum from the weathering of bedrock, through the movement of particles by the force of gravity (mass wasting) or movement caused by a transporting agent such as water or wind (surface erosion), to the eventual deposition of

Corresponding author. Tel: +1-360-753-7691;

fax +1-360-956-2346

E-mail address: swondzell@fs.fed.us

particles in ocean basins. Wind erosion is uncommon in most forested areas, so we focus on types of surface erosion resulting from overland flow of water, which includes uniformly distributed sheet erosion, rill and interrill erosion, and gully erosion on hillslopes and both channel incision and bank-cutting in stream channels. We also focus on hillslope mass-wasting processes, including ravel, soil creep, deep-seated earth flows, debris slides and debris flows.

2.1 Surface Erosion

Because surface erosion usually requires overland flow of water, its occurrence is dependent upon the factors that control runoff generation, namely, soils, vegetation and water input (precipitation or snowmelt). Two mechanisms can generate overland flow: 1) saturation of the soil to the surface, and 2) water input rates exceeding infiltration rates. The role of saturationexcess overland flow in erosional processes appears to be little studied. In contrast, infiltration-excess overland flow (or Hortonian flow) has been well studied. It is the dominant mechanism driving erosion in arid and semiarid regions, but is relatively uncommon under wetter climatic regimes. The dense overstory and understory vegetation found in most forests, combined with well developed litter layers, protect the soil surface from rain splash. Also, the litter layer can store substantial amounts of water and thereby regulate the rate at which infiltrating rainwater reaches the mineral soil surface (Martin and Moody 2001). Further, many forest soils are well structured which also promotes rapid infiltration. Thus, infiltration-excess overland flow is rare from undisturbed forest soils, and is usually confined to local areas (Harr 1979, Troendle and Leaf 1980).

There is great variation in forest types from the coastal Pacific Northwest through the mountains of the Interior Columbia Basin, to the Northern Rocky Mountains. In general, differences in forest type reflect differences in climate. Maritime climates in the Pacific Northwest tend to be wetter than interior climates and are dominated by rain and rain-on-snow precipitation regimes. The Interior region is snow-melt dominated, but also receives intense summer thunderstorms. Throughout both regions, precipitation increases with elevation. In the wet maritime climates of the Pacific Northwest region, however, precipitation is sufficient to support dense forested vegetation from sea level to the high-elevation tree line. In the Interior region, elevational differences in temperature, precipitation, and evapotranspiration create steep gradients in plant-available moisture that result in striking differences in forest cover. Forests tend to be restricted to wetter mountainous areas, and in many places, mountain foothills and even lower elevation mountain ranges may be too dry to support forests. At the lower elevational limit of forests, especially on south

facing hillslopes, trees are often widely spaced, total vegetatitive cover can be low, and litter layers may be poorly developed. Under these conditions, infiltration-excess overland flow may occur regularly (Wilcox et al. 2001). Coastal rainforests of the Pacific Northwest represent the other extreme, where annual precipitation can exceed 300 cm but overland flow is generally not observed from undisturbed forest soils.

Many studies have documented dramatic increases in surface erosion following wildfire (Helvey 1980, Meyer and Wells 1997, Robichaud and Brown 1999, Cannon et al. 2001, Meyer et al. 2001, Moody and Martin 2001a). Many factors can account for accelerated surface erosion, and the exact blend of mechanisms contributing to increased erosion changes among locations and with the sequence of post-fire meteorological events. The primary factors are the availability of readily erodible sediment and changes in soil infiltration rates. Easily erodible sediment is exposed to surface erosion when fire removes ground-covering vegetation, litter and organic layers that previously protected it from detachment (Johansen et al. 2001). Loss of soil structure from intense heating, combustion of soil organic matter, and soil drying can lead to decreased cohesiveness of surface soil aggregates, which are then more readily eroded. Additionally, burning of logs or other organic obstructions on hillslopes can liberate previously stored sediment to surface erosion. Finally, physical disturbances of the soil, such as windthrow or disturbance by animal activity (Swanson 1981) all contribute to increasing the amount of sediment available to be eroded.

Overland flow can physically detach and transport sediment and is the dominant mechanism of surface erosion after wildfire. The immediate causal factors most changed by wildfire are the soil, litter, and vegetative properties that determine infiltration rates. Ground-cover vegetation, litter, and soil organic horizons all protect the mineral soil from rain-drop impacts that can dislodge soil particles, mobilizing sediment to be eroded. Rain splash can also disrupt and possibly even compact the soil (Meyer and Wells 1997); fine sediment dislodged by rain splash can clog soil pores causing surface sealing (Swanson 1981, Wells et al. 1979, Martin and Moody 2001). Surface sealing is further accentuated immediately after wildfire when organic matter binding soil aggregates has been combusted, so that aggregates easily disintegrate with physical disturbance, and when ash on the soil surface provides an abundance of fines (Swanson 1981. Meyer and Wells 1997, Cannon et al. 2001). Finally, heating soil organic matter can form hydrophobic compounds that coat soil particles and create a waterrepellent layer. Some studies have shown that formation of hydrophobic compounds may be dependent on the type of vegetation, the antecedent soil-moisture content, and soil texture (Wells et al. 1979, McNabb and Swanson 1990, DeBano et al. 1998, Robichaud and Hungerford

2000, Huffman et al. 2001). In all cases, however, the formation of hydrophobic compounds depends on the soil temperature attained during a fire. Thus, the presence of a water repellent layer, and its depth in the soil profile is largely determined by fire behavior, fire severity and soil temperature gradients during a fire (DeBano 2000, Robichaud and Hungerford 2000).

Reduced rates of infiltration are usually observed after severe fires (Fig. 1) in both the Pacific Northwest region (Swanson 1981, McNabb et al. 1989, Johnson and Beschta 1980), the Interior region (Robichaud 2000), and in other regions (Johansen et al. 2001, Martin and Moody 2001, Wohlgemuth et al. 2001, Benavides-Solorio and MacDonald 2001). However, the relative importance of the different physical mechanisms potentially reducing infiltration rates is not known. Further, their relative importance probably varies in time and place, with fire intensity and duration, and with time since the last fire. What is clear is that reduced infiltration after severe wildfire can contribute to increased overland flow and accelerated erosion (Elliott and Parker 2001).

Reductions in infiltration rates reduce the threshold precipitation intensity at which overland flow occurs. In recently burned forests, precipitation intensities with recurrence intervals of 5 yr or less can exceed infiltration rates, whereas precipitation intense enough to exceed infiltration rates in unburned forested areas reoccurs approximately once every 30 years (Fig. 1). Although accelerated erosion from overland flow on burned slopes is well documented for the Interior region and other areas dominated by continental climates (Megahan et al. 1995, Meyer and Wells 1997, Cannon et al. 2001, Martin and Moody 2001, Meyer et al. 2001, Moody 2001, Moody and Martin 2001a) it has not been documented from burned-forest areas in the maritime-climate dominated Pacific Northwest region.

The regional differences in the occurrence of infiltrationexcess overland flow are typically attributed to climatic differences. Summer rainfall in the continental climatedominated areas of the Interior region primarily results from thunderstorms. These storms occasionally generate intense rainfall, driving infiltration-excess overland flow, especially from burned areas where infiltration rates are reduced. In contrast, many authors have suggested that the maritime climates of the Pacific Northwest region are characterized by long-duration, low-intensity rainfall, so that infiltration rates are seldom exceeded, even after intense wildfires (Swanson 1981, Beschta 1990, Wondzell 2001). This is not supported by our analysis of regional differences in rainfall intensity (Fig. 1). Rainfall intensities in excess of expected infiltration rates appear more common in the maritime climate of the Pacific Northwest region than in thunderstorm-dominated continental climate of the Interior region. Further, while measured infiltration rates are highly variable, there is no

evidence to suggest that infiltration rates are inherently higher in soils of the Pacific Northwest region than in the Interior region, nor are post-fire changes in infiltration rates notably different among the regions (Fig. 1). These data raise interesting questions as to why infiltrationexcess overland flow and attendant erosion have not been documented in the Coastal and Cascade Mountains of the Pacific Northwest.

Rapid recovery of fire-caused reductions in infiltration rates, high antecedent soil moisture, and rapid rates of vegetative regrowth after fires might all explain why postfire, infiltration-excess overland flow has not been reported in the Pacific Northwest region. First, the soils of many unburned forested areas are hydrophobic when dry (Benevides-Solorio and MacDonald 2001, Huffman et al. 2001. However, hydrophobicity is not evident in these soils once moisture content exceeds 12% to 25% (Huffman et al. 2001). Secondly, hydrophobicity tends to increase following fire. Fire-caused hydrophobic layers can persist in the soils of some forest types for long periods. For example, hydrophobic layers in the soils of dry pine forests can persist for months to years (Dyrness 1976, Huffman et al. 2001), and after intense wildfires, reduced infiltration rates can persist for as long as 6 years (Dyrness 1976). In contrast, McNabb et al. (1989) showed rapid loss of hydrophobicity and rapid recovery of infiltration rates after prescribed fires in Coastal mountains of southern Oregon.

In the Interior region, where summer rainfall is from thunderstorms, it is likely that intense rain will fall on dry soils. In the Pacific Northwest region, in contrast, intense rain is much more likely to fall on wet soils. Firstly, thunderstorms occur on 20-30 days during the summer in the Interior region, where as thunderstorms occur less than 5 days per year in Pacific Northwest region (Miller et al. 1963). Secondly, the rainy season is long in the Coastal and western Cascade Mountains, driven by frequent frontal storms off the northern Pacific. These storms do bring intense rainfall to the Pacific Northwest, but these usually occur as an intense storm cell imbedded in the larger, frontal storm.

The prevalence of low-intensity rainfall, relatively high soil moisture, and rapid recovery of ground-covering vegetation after fires might all restrict the window of time after burning during which soils are at risk of accelerated erosion from infiltration-excess overland flow in maritime-climate dominated areas of the Pacific Northwest region. These factors, then, could substantially reduce the probability that surface erosion will occur, and given the relative rarity of severe fires in the Coastal and Cascade mountains in recent decades, perhaps it is not surprising that infiltration-excess overland flow and widescale surface erosion have not been documented within the region.



Fig 1. Soil infiltration rates in burned and unburned forest areas (top panel) and rainfall intensity (bottom panel). Infiltration data is from ¹Robichaud 2000, ²Martin and Moody 2001a, ³McNabb et al. 1989, and ⁴Johnson and Beschta 1981. Abreviations denote dominant tree species in the forests at each study site. PSME = *Pseudotsuga menzesii* (Douglas fir); PICO = *Pinus contortus* (lodgepole pine); PIPO = *Pinus ponderosa* (ponderosa pine). Rainfall data from NOAA 15-minute precipitation records for the Quinault Ranger Station, WA (Coastal Pacific Northwest); Ukiah, OR (Interior Pacific Northwest); and McCall, ID (Northern Rocky Mountains).

2.2 Mass Wasting

Mass-wasting is most common in the steep topography of mountainous areas, but can occur anywhere geomorphic processes create steeply sloping landforms, including steep valley-side slopes or cut banks above active river channels. Geologists and geomorphologists recognize many unique classes of mass wasting events and have developed a systematic classification scheme and naming conventions for mass-wasting processes (see Varnes 1978), however, the physical basis for distinguishing among types of mass-wasting and erosional transport events is complex. Many mass-wasting events include several uniquely defined processes linked in a sequence along a stream network that Nakamura et al. (2000) called a disturbance cascade. For example, debris slides commonly create debris flows which form debris jams. If debris jams impound water and subsequently

burst, they can release flood surges. Further, flow properties of sediment in transport may change substantially within a single event, ranging across the continuum from debris flows, to hyperconcentrated flows, to sediment-laden water floods (Costa 1988, Grant Meyer, personal communitcation). Although consideration of all possible mass wasting and erosional transport processes is beyond the scope of this paper, it is important to recognize that different types of processes may have different effects on stream channel morphology, and thus have variable influence on the habitat of fish and other aquatic organisms. For this paper, however, we focus on a few basic types of mass-wasting processes, namely, ravel, debris slides, and debris flows. Further, we follow Nakamura et al. (2000), and generically refer to the shallow sliding of rock, sediment, and soil on hillslopes as debris slides. Debris slides can reach stream channels, and the down-channel flow of that rock, sediment, and soil we call debris flows.

Severe fires clearly increase the frequency and magnitude of a variety of episodic mass-wasting events. Post-fire debris slides and debris flows are the most frequently studied post-fire mass-wasting processes. Other mass-wasting processes affected by fire include soil creep and deep-seated earth flows (Swanson 1981), both of which are set in motion by soil saturation. Loss of forest canopies decreases evapotranspiration and can result in more frequent or longer periods of soil saturation. Although the physical cause-effect relationship is clear, we do not know of any studies providing empirical evidence to confirm that view. However, forest removal by logging has resulted in wetter soil conditions and generally higher creep rates (Gray 1977). Increased peak flows may also occur following severe wildfire (Helvey 1973, Cheng 1980, Elliot and Parker 2001, Moody and Martin 2001b), which in turn can cause bank erosion and bank-side slides that increase delivery of wood and sediment to channels.

2.2.1 Ravel:

Ravel (often called dry ravel) is the rapid downhill movement of individual particles and can include both organic and inorganic materials of various sizes (Swanson 1981). Ravel occurs preferentially on steep to very steep slopes. Mersereau and Dyrness (1972) found 4 times more ravel from 80% slopes than from 60% slopes. Also, ravel is much greater in noncohesive soils. Many soils lose cohesiveness on drying, so that ravel preferentially occurs during the dry season, and from exposed, southfacing hillslopes. Vegetation tends to stabilize the soil surface, so that little ravel occurs, even on steep, southfacing slopes, if vegetation cover exceeds 50% to 75% (Mersereau and Dyrness 1972).

Ravel is unique, in that it is the only mass-wasting process accelerated by wildfire that occurs independently of post-fire storm events. Severe fire removes litter, duff and vegetation that stabilize the soil; heating combusts soil organic matter that binds soil aggregates and dries the soil decreasing soil cohesion; and fire consumes logs and other organic barriers that store sediment thereby making more sediment available to be moved by ravel. Fire effects on ravel may be short lived. One study from the Oregon Coast Range found that ravel occurring in the first 24 hours after burning accounted for approximately 2/3 of the total ravel measured in the first year after burning (Bennett 1982). Another study in the western Cascade Mountains showed that rates of ravel were reduced to near zero by the second growing season after prescribed burning which was attributed to rapid recovery of vegetation (Mersereau and Dyrness 1972). Accelerated rates of ravel might be expected to persist much longer wherever post-fire vegetation recovery is slow, for

example, in low-elevation, dry forest types growing on south-facing slopes in the Interior region, especially on noncohesive soils derived from granitic parent materials. Megahan et al. (1995) showed that accelerated erosion rates persisted for at least 10 years on south-facing slopes following helicopter logging and prescribed burning, probably because low water availability limited the rate of vegetative recovery after burning. In contast, near complete recovery of accelerated sediment yields was observed on north facing slopes by 3 years after burning. However, these were watershed-scale studies (Megahan et al. 1995), so the relative contribution of ravel and surface erosion to the sediment budgets cannot be differentiated. Ravel can be substantial after severe fires and can contribute to sediment loading of channels adjacent to steep slopes, but in many cases, raveled sediment will only be transported short distances before being captured in storage locations.

2.2.2 Debris Flows:

Debris flows can be initiated in two ways – either from surface runoff or from debris slides. Because overland flow seldom occurs in the Pacific Northwest region, runoff-initiated debris flows have not yet been documented within the region. In contrast, both types of initiation events have been documented for debris flows in the Interior region, and else where throughout western North America. Numerous studies have documented increased frequency of debris flows following severe wildfire.

2.2.2.1 Runoff-initiated debris flows

Debris flows can be initiated by overland flows of water, although there is some debate as to the exact mechanism through which such debris flows are generated. Meyer and Wells (1997) and Cannon et al. (2001) suggest that runoff-initiated debris flows occur when surface runoff entrains fine sediment over large areas and converges to begin carving small rills and larger gullies, eventually entraining sufficient sediment to form debris flows on hillslopes or high in the channel network. In other cases, sediment-laden water floods in steep headwater channels must entrain additional sediment from channel and bank erosion to transition to debris flows (Meyer and Wells 1997). These events are often referred to as bulking flows. Alternatively, Wells (1987) and DeBano (2000) suggested that tiny debris slides from saturated soils above hydrophobic layers a few centimeters deep create debris-flows, which in turn carve the network of rills and small gullies commonly observed in the initiation zone of these events. However, neither Meyer and Wells (1997) nor Cannon et al. (2001) have seen evidence of tiny debris slides at their study sites. Regardless the specific sequence of events that initiates

these debris flows, it is clear that they are not caused by *en masse* release of sediment from large debris slides.

Runoff-initiated debris flows are relatively common. and have been observed in a variety of environments including the northern Rockies (Meyer and Wells 1997). the southern Rocky Mountains (Cannon et al. 2001), the interior Northwest (William Russell, Oregon State University, personal communication), and California (Wells 1987), but not in the Coastal or western Cascade mountains of the Pacific Northwest. Because these debris flows result from surface runoff, they are ultimately controlled by the same suite of factors that control surface erosion, namely rainfall intensity and soil-infiltration rates. Also, they respond similarly to fire-induced changes in infiltration rates, with one important additional factor. Both Meyer and Wells (1997) and Cannon et al. (2001) suggest that the abundance of fine sediment and ash on the surface of recently burned soils is critical to generating debris-flow conditions.

2.2.2.2 Debris slides and debris-slide-initiated debris flows

Debris slides occur when a large mass of sediment, often 100s of cubic meters in size, moves en masse on steep hillslopes. Debris slides tend to be associated with major storm events and floods (Megahan et al. 1978, Rapp et al. 1991, McClelland et al. 1997). In many cases, major storm systems are large enough to cause widespread impact from debris slides and debris flows over large regions. For example, numerous slides and debris flows were recorded from the Pacific coast to central and northern Idaho during the winter of 1996, a year marked by 50- to 100-year return interval floods (McClelland et al. 1997, Hofmeister 2000, Nakamura et al. 2000. Meyer et al. 2001). Of course debris slides and debris flows occur during smaller magnitude storms, but in these cases, debris slides are relatively less frequent and are not as widespread.

Swanston (1971) applied soil mechanic theory to describe debris slide initiation. This theory predicts that debris slides occur if shear stresses equal or exceed shear strength (the combined resistance to movement provided by friction against the shear plane and internal cohesion of the soil). The balance between shear stress and shear strength is primarily a result of slope steepness and both the type and thickness of sediment. However, shear stresses and shear strength are also influenced by a variety of external processes. These include increased loading of the soil mass through increases in water content during storms, or from sediment deposition caused by ravel, soil creep, earth flow, or surface erosion; the loss of physical support, for example by bank erosion undercutting steep slopes above a river channel; decreased frictional resistance caused by increased pore-water pressures when soils become saturated; and finally, decreases in internal

cohesion caused by changes in soil moisture content and loss of mechanical cohesion provided by roots (Swanston 1971, Swanson 1981).

Fire indirectly influences the balance between shear stresses and shear strength. For example, fire accelerates rates of ravel and surface erosion, depositing sediment in hillslope hollows, a common initiation point for debris slides (Dietrich et al. 1982). Fire can also lead to increased peak flows and may therefore contribute to accelerated bank erosion with a concurrent increase in rates of bank-side sliding. However, increased soil-water content and decreased root strength are the most important factors leading to accelerated rates of debris sliding after fire (Swanston 1971, Swanson 1981).

The close association of debris slides with extreme storm events results from the relationship between pore water pressures and the rate at which water is added to the soil (Swanston 1971). Thus, discounting other factors, the relative likelihood of debris slide occurrence should be highly correlated with the probability of occurrence of extreme storms. Regional trends in 50 year returnfrequency storms, show that 24-hour precipitation totals range from 17 cm to more than 25 cm in the Pacific Northwest region, but decrease to only 7 to 13 cm east of the crest of the Cascades in the Interior region (NOAA undated). These storm effects on debris slide occurrence are further accentuated by rain-on-snow events that can greatly increase the amount of water flowing into the soil. Winter temperatures tend to be mild west of the crest of the Cascade mountains, creating a transitional snow zone. Snow levels may drop to 500 m in elevation, or less. during particularly cold frontal storms. Warmer storms may bring rainfall to 1200 m in elevation, or even much higher. Thus, between elevations of approximately 500 to 1000 m, deep snow can accumulate, but the snow pack tends to be warm and very wet. These ripe snow packs melt rapidly during major warm storms, dramatically increasing the amounts of water reaching the soil. In contrast, the snow pack tends to be colder and drier in the more continental climatic regime of the Interior region. Therefore, it takes longer for the snow pack to begin to melt and release water to the underlying soil when warm Pacific frontal systems bring rain to the mountains of the interior during the winter. Thus, rain-on-snow events should be relatively less common in the interior. When they do occur, however, they are often associated with widespread occurrence of debris slides (Megahan et al. 1978). The Interior region is characterized by snow-melt dominated hydrologic systems. Years in which the spring season is long and cool lead to slow melting of the snow pack so that few debris slides occur. However, a rapid shift to hot weather in the spring can lead to rapid melting of the snow pack and trigger debris slides (Megahan et al. 1978, Helvey 1980). Spring snow melt has not been identified as an important mechanism triggering debris slides in the Pacific Northwest region.

Storm effects on debris slides are greatly affected by loss of the forest canopy caused either by stand-replacing wildfire or clearcut harvesting. First, evapotranspiration is decreased so that soils remain wetter over longer periods (Swanston 1971, Klock and Helvey 1976, Helvey 1980, Swanson 1981, McNabb and Swanson 1990). Consequently, the threshold of storm magnitude needed to bring the soil to saturation and trigger debris slides can be reduced after severe wildfire. Loss of the forest canopy also accentuates the effect of rain-on-snow events. Energy budgets of snow melt during rain-on-snow events shows that the relatively "warm" rain provides little energy to melt snow. Rather, the primary source of energy to melt snow is the condensation of water vapor onto the snow pack (Harr 1981, Swanson 1981, McNabb and Swanson 1990). Dense forest canopies shelter the snow surface from strong winds. After a stand replacing fire, however, wind reaches the surface of the snow pack where vapor from the warm, humid air condenses directly onto the snow pack so that prolonged storm events can melt substantial amounts of snow.

Decreases in internal cohesion caused by loss of mechanical cohesion as roots of fire-killed trees decompose can also decrease the effective soil strength, making slopes more susceptible to debris sliding (Swanston 1971, Swanson 1981, McNabb and Swanson 1990). Several studies show an apparent increase in debris slide occurrence 5 to 10 years after severe wildfire or clearcut harvesting, and suggest this pattern would be consistent with temporal trends in loss of mechanical cohesion from decomposing roots (Megahan et al. 1978).

The movement and transport of sediment from debris slides may follow one of several different trajectories (Nakamura et al. 2000). The initial failure and movement of sediment may occur as a block of soil and sediment that remains relatively intact, and moves only a short distance before coming to rest. Alternatively, initial failure and movement may lead to rapid disaggregation of the slide mass and formation of a debris slide. On concave slopes, debris slides may be deposited on lower angled hillslopes below the initiation point. Alternatively, the debris slide may continue down slope, eventually reaching the channel network (Nakamura et al. 2000).

Once debris slides reach the channel network, they can be deposited as debris jams or can be mobilized into debris flows. Debris flows may stall in lower-gradient stream reaches or at tributary junctions with larger streams, especially if the channel junction occurs at oblique angles, or if the tributary crosses a large, low angled alluvial fan or a wide floodplain developed in the mainstem valley floor (Nakamura et al. 2000). Alternatively, debris flows may continue long distances down relatively large streams (Wondzell and Swanson 1999). Wherever debris flows finally stop, they typically construct large jams of sediment and wood which often block the stream channel and create zones of sediment deposition immediately upstream (Montgomery et al. 1996, Wondzell and Swanson 1999, Benda *this volume*). During major floods, debris jams can impound water, and in some cases, may fail catastrophically releasing flood surges downstream (Nakamura et al. 2000).

2.2.2.3 Contrasting runoff-initiated vs. debris-slideinitiated debris flows

In areas where both runoff-initiated and debris-slide initiated debris flows occur, it is difficult to assess the relative importance of the different initiation sequences in sediment budgets. The difference in occurrence should depend on the relative probability of debris slides and the relative frequency of overland-flow generation. However, even in areas with known, high debris-slide hazard, there is great variation among geologic parent materials and among landforms in the relative susceptibility of slopes to debris sliding, regardless of the degree of disturbance (Swanson and Dyrness 1975, McClelland et al. 1999). Similarly, the likelihood of surface runoff is dependent on climatic, soil and vegetation factors, and both soil properties and vegetation change with time following fire. Consequently, great local and regional variation should be expected in the relative frequencies of the two debris-flow initiation mechanisms.

Meyer et al. (2001) suggested a general hypothesis to explain differences in the timing of runoff-initiated and debris-slide initiated debris flows. They suggested that runoff-initiated debris flows tend to occur within one or two years after a severe fire while debris-slide initiated debris flows will tend to occur some 5 to 10 years after a fire. In most places, fire-induced water repellency is short lived because hydrophobic compounds break down in a few years. Similarly, stripping of fines from the soil surface in previous erosional events, and both compaction and increased cohesion of the surface soil layer reduce the influence of fine sediment and ash on infiltration rates. Additionally, recovery of ground-cover vegetation rapidly stabilizes the soil surface. All these processes reduce the likelihood of overland flow and surface erosion and also reduce the availability of fine sediment needed to generate debris flows (Meyer and Wells 1997). During this same period, however, roots from fire-killed trees are decomposing, and while tree seedlings may be reestablishing in the burned areas, it will be many years before they regrow extensive root networks. Thus, mechanical cohesion should reach a minimum between 5 and 10 years after wildfire. In both cases, however, extreme climatic events are needed to trigger these debris flows. The high variability in the timing of extreme climatic events often prevents such clear-cut sequences in the timing of episodic erosional events.

3. Discussion

Erosional processes following wildfire are distinctly different between the forested landscapes of the Interior and Pacific Northwest regions. Because erosion is controlled by a variety of factors relating to soils, geology, topography, vegetation and climate, variability in erosional processes within a region may be as large as variability between regions. However, in reviewing the available literature several major differences in erosional processes are apparent between these two regions.

Surface erosion from infiltration-excess overland flow is a dominant response after wildfire in the Interior region, but has not been documented in the Pacific Northwest region. The likelihood of surface erosion from overland flow is a function of the probability of intense storm occurrence and the surface soil conditions regulating infiltration. Measured post-fire infiltration rates and precipitation intensities (Fig. 1) suggest that surface runoff from infiltration-excess overland flow could occur in the Coastal and Cascade Mountains of the Pacific Northwest after fires. We hypothesize that the frequency of high intensity convective storms during the summer in the Interior region increases the probability of rainfall on dry soils and when combined with generally less rapid vegetation recovery substantially increases the likelihood of overland flow and surface erosion following fire.

In both the Pacific Northwest region and the Interior region, increases in debris slides and debris-slide initiated debris flows can occur following wildfire because of increased soil-water content and decreased root strength. Debris slides and related debris flows are more probable in the Pacific Northwest region due to greater annual precipitation, larger storms (e.g. 24 hour precipitation totals for 50-yr return-frequency events) and more widespread occurrence of rain-on-snow events. In addition, the reduction in evapotranspiration and the resulting increase in soil water following vegetation removal generally tend to be larger for regions with larger annual precipitation. Also, forest removal in the rain-onsnow zone in much of the Pacific Northwest region may cause increases in latent heat inputs to the snowpack during rain events such that water input rates to the soils are greatly enhanced. Loss of shear strength of the soil over time after a severe fire, as tree roots decay, typically results in increases in debris slide occurrence 5-10 years post-fire.

In the Interior region, increases in runoff-initiated debris flows can also occur following wildfire; however, these have not been observed in the Pacific Northwest region. In the Interior region, the mechanisms that generate overland flow can lead to debris flows if discharge and available sediment are sufficient. Thus, the abundance of fine sediment and ash on the soil surface of recently burned areas may be critical in generating debris flows (Meyer and Wells 1997, Cannon et al. 2001). The likelihood of runoff-initiated debris flows is a function of the probability of intense storm occurrence, the surface soil conditions regulating infiltration and the abundance of surface fines.

In the Interior region the time frame of susceptibility to accelerated erosion following fire is considerably longer than in the Pacific Northwest region. Immediately following the fire and for some time period thereafter, sites are susceptible to overland runoff and related surface erosion and debris flow occurrence. Areas are also susceptible to debris slide and related debris flow activity about five to ten years following fire. This later window of susceptibility appears to be the primary time frame for accelerated fire-related erosion in the Pacific Northwest region.

Our review of the literature suggests that severe wildland fires affect hillslope erosion and stream sedimentation in forested watersheds. However, this conclusion probably presents an unbalanced assessment of the overall effect of fire on erosion and sedimentation rates. Most of the studies in the literature, and therefore most of the studies cited in this review, examine the effects of either 1) severe wildfire followed by large to extreme storms that generate episodic erosion, or 2) clearcut harvesting followed by "prescribed" burning of residual logging slash (especially for studies of fire effects in the Pacific Northwest region). Consequently, we know little about effects of moderate to low severity wildfire on erosion. Also, our knowledge about fire effects is confounded by the effects of other landuse practices, especially logging and road building. Clearly, more research is needed to better understand the effects of moderate to low severity fires, including prescribed fires, on erosional processes in forested watersheds.

Our knowledge of the role of massive, episodic inputs of sediment to streams is poor. Firstly, because they are episodic, it is difficult to make good estimates of the relative contribution of episodic inputs to sediment budgets. Erosion rates have been measured in a variety of studies, from small-plot studies under either ambient climate or artificial precipitation, to small watershed sediment budgets, to the use of radionuclide tracers (cesium-137) deposited in the 1950s and 1960s (Coppinger et al 1991) from testing of nuclear bombs, to the use of cosmogenic isotopes. Different types of studies provide estimates of erosion rates integrated over different time periods-from single storms in some plotscale studies to thousands of years in cosmogenic isotope studies-and consequently result in different estimates of erosion rates. Erosion estimates derived from short-term, small-watershed studies often considerably underestimate long-term rates of erosion because these studies typically miss the infrequent but large, episodic events such as debris slides and debris flows (Kirchner et al. 2001). Alternatively, extending erosion rates measured in smallplot studies to watersheds or regions may overestimate

erosion rates because much sediment eroded from hillslopes is only transported a short distance before being redeposited. Sediment eroded from hillslopes may also be deposited on alluvial fans or floodplains where it can be stored for 100s to 1000s of years (Meyer et al. 1995, Clayton and Megahan 1997, Trimble and Crosson 2000, Moody and Martin 2001a). These problems make it difficult to estimate erosion rates and determine how they will change in response to a specific disturbance such as fire. More research is needed to understand better the links and feedbacks between fire, surface erosion, and episodic mass-wasting events.

The effectiveness of different types of erosional, mass-wasting, and sediment transport events in shaping stream channels should be expected to differ, and therefore should influence ecological functions in stream ecosystems in different ways. For example, the relative proportion of water and sediment, and the abundance of fine sediment, determine the flow properties of eroded materials while in transport (Costa 1988). Fluid-like flows of sediment-laden water floods do not generate high shear stresses (Costa 1988), and as a consequence, they should have relatively less impact on channel morphology than do debris flows. Debris flows are much less fluid, moving as a viscoplastic mass (Costa 1988) that can dramatically reshape channel morphology (Wondzell and Swanson 1999, Benda, this issue). Also, debris flows are more competent to transport large boulders and logs and deposit them in stream channels, thus adding physical structure to stream channels. The morphology of channels shaped by debris flows, combined with the coarse sediments and wood delivered to stream channels may be important for maintaining long-term habitat diversity and suitable spawning gravels in some stream systems (Swanston 1991, Reeves et al. 1995). In contrast, surface erosion from sheet flow or rill networks are more likely to deliver only fine-textured sediment and fine-particulate organic matter to streams. The input of fine sediment and its subsequent movement downstream will have different effects on aquatic habitat than will large particles delivered by debris flows. The links between some erosional and mass-wasting processes and channel morphology are well studied. The links between some channel morphologic features and a variety of ecosystem, community and population responses in aquatic ecosystems are also well studied. However, we still know relatively little about how different types of erosional and mass-wasting events will influence stream ecosystem processes and the stream habitats required by aquatic species.

Acknowledgements

We thank Tom Badger, Pete Bisson, Wendy Gerstel, Charlie Luce, Grant Meyer, and Will Russell for discussions that improved this manuscript. We also thank two anonymous reviewers for helpful comments.

References

- Benavides-Solorio, J. and MacDonald, L. H. 2001. Postfire runoff and erosion from simulated rainfall on small plots, Colorado Front Range. Hydrol. Process. 15, 2931-2952
- Benda, L.E. (*this issue*). The role of fire and punctuated sediment and wood supplies on channel and valley morphologies and aquatic habitats. Submitted to For. Ecol. Manage.
- Bennett, K.A. 1982. Effects of slash burning on surface erosion rates in the Oregon Coast Range. M.S. Thesis, Oregon State University, Corvallis.
- Berris, S. N. and Harr, R. D. 1987. Comparative snow accumulation and melt during rainfall in forested and clear-cut plots in the western Cascades of Oregon. Water Resour. Res. 23, 135-142
- Beschta, R. L. 1990. Effects of fire on water quantity and quality. In: J. D. Walsad, S. R. Radosevich and D. V. Sandberg (eds.) Natural and Prescribed Fire in the Pacific Northwest Forests. Corvallis OR., Oregon State University Press. pp. 219-231.
- Cannon, S. H., Bigio, E. R., and Mine, E. 2001. A process for fire-related debris flow initiation, Cerro Grande fire, New Mexico. Hydrol. Process.15, 3011-3023
- Cheng, J. D. 1980. Hydrologic effects of a severe forest fire. In: Proceedings of Symposium on Watershed Management. American Society of Civil Engineers, New York, pp. 240-251.
- Clayton, J. L. and Megahan, W. F. 1997. Natural erosion rates and their prediction in the Idaho batholith. J. Am. Water Resour. Assoc. 33, 689-703
- Coppinger, K. D., Reiners, W. A., Burke, I. C., and Olson, R. K. 1991. Net erosion on a sagebrush steppe landscape as determined by Cesium-137 distribution. Soil Sci. Soc. Am. J. 55, 254-258
- Costa, J. E. 1988. Rheologic, geomorphoic, and sedimentologic differentiation of water floods, hyperconcentrated flows, and debris flows. In: V.R. Baker, R. C. Kochel and P. C. Patton (eds.) Flood Geomorphology. pp. 113-122.
- Dietrich, W. E., Dunne, T., Humphrey, N., and Reid, L. M. 1982. Construction of sediment budgets for drainage basins. In: F. J. Swanson, R. J. Janda, T. Dunne, and D. N. Swanston (eds.), Sediment budgets and routing in forested drainage basins. Gen. Tech. Rep. USDA For. Ser. Pac. Northwest For. Range Expt. Stn. PNW-141. Portland, Oregon. pp. 5-23.
- DeBano, L. F., 2000. The role of fire and soil heating on water repellency in wildland environments: a review. J. Hydrol. 231-232,195-206.
- DeBano, L. F., Neary, D. G., and Ffolliott, P. F. 1998. Physical Soil System. In: DeBano, L. F., Neary, D.

G., Ffolliott, P. F. Fire's Effects on Ecosystems. John Wiley & Son's, New York. pp. 84-102.

- Dyrness, C. T. 1976. Effect of wildfire on soil wettability in the high Cascades of Oregon. Res. Pap. USDA For. Ser. Pac. Northwest For. Range Expt. Stn. PNW-202. Portland Oregon..
- Elliott, J. G. and Parker, R. S. 2001. Developing a postfire flood chronology and recurrence probability from alluvial stratigraphy in the Buffalo Creek watershed, Colorado, USA. Hydrol. Process. 15, 3039-3051
- Gray, D. H. 1977. Creep movement and soil moisture stress in forested vs. cutover slopes: results of field studies (final report). National Science Foundation, Grant No. ENG 74-02427, 141p.
- Gresswell, R. E. 1999. Fire and aquatic ecosystems in forested biomes of North America. Trans. Am. Fish. Soc. 128, 193-221
- Harr, R. D. 1979. Effects of timber harvesting on streamflow in the rain-dominated portion of the Pacific Northwest. In: Proceedings of workshop on scheduling timber harvest for hydrologic concerns. USDA For. Ser. Pac. Northwest Region Pac. Northwest For. Range Expt. Stn. Portland, Oregon. pp. 2-45.
- Harr, R. D. 1981. Some characteristics and consequences of snowmelt during rainfall in western Oregon. J. Hydrol. 53,277-304
- Helvey, J. D. 1973. Watershed behavior after forest fire in Washington. In: Proceedings of the Irrigation and Drainage Division Specialty Conference, American Society of Civil Engineers, April 22-24, 1973, Fort Collins, CO. pp. 403-422.
- Helvey, J. D. 1980. Effects of a north-central Washington wildfire on runoff and sediment production. Water Resour. Bull. 16, 627-634
- Hofmeister, R. J. 2000. Slope failures in Oregon: GIS inventory for three 1996/97 storm events. Special Paper 34, Oregon Department of Geology and Mineral Industries.
- Huffman, E. L., MacDonald, L. H., and Stednick, J. D. 2001. Strength and persistence of fire-induced soil hydrophobicity under ponderosa and lodgepole pine, Colorado Front Range. Hydrol. Process. 15, 2877-2892
- Johnson, M. G. and Beschta, R. L. 1980. Logging, infiltration capacity, and surface erodibility in western Oregon. J. For. June,334-337
- Johnson, M. G. and Beschta, R. L. 1981. Seasonal variation of infiltration capacities of soils in western Oregon. Res. Note USDA For. Serv. Pac. Northwest For. Range Expt. Stn. PNW-373. Portland, Oregon.
- Johansen, M. P., Hakonson, T. E., and Breshears, D. D. 2001. Post-fire runoff and erosion from rainfall simulation: contrasting forests with shrublands and grasslands. Hydrol. Process. 15, 2953-2965

- Kirchner, J. W., Finkel, R. C., Riebe, C. S., Granger, D. E., Clayton, J. L., King, J. G., and Megahan, W. F. 2001. Mountain erosion over 10-year, 10,000-year, and 10,000,000-year timescales. Geology 29, 591-594
- Klock, G. O. and Helvey, J. D. 1976. Soil-water trends following wildfire on the Entiat Experimental Forest. In: Annual Proceedings of Tall Timbers Fire Ecology Conference No. 15. USDA For. Serv. Tall Timbers Res. Stn. Tallahassee Florida, pp. 193-200.
- Martin, D. A. and Moody, J. A. 2001. Comparison of soil infiltration rates in burned and unburned mountainous watersheds. Hydrol. Process. 15, 2893-2903
- McClelland, D. E., Foltz, R. B., Wilson, W. D., Cundy, T., Heinemann, R., Saurbier, J., and Schuster, R. L. 1997. Assessment of the 1995 & 1996 floods and landslides on the Clearwater National Forest. Part I: Landslide Assessment. A report to the Regional Forester. USDA For. Serv. Northern Region. Missoula, Montana.
- McClelland, D. E., Foltz, R. B., Falter, C. M., Wilson, W. D., Cundy, T., Schuster, R. L., Saurbier, J., Rabe, C., and Heinemann, R. 1999. Relative effects on a low-volume road system of landslides resulting from episodic storms in nothern Idaho. Transp. Res. Rec. 1652, 235-243
- McNabb, D. H., Gaweda, F., and Froehlich, H. A. 1989. Infiltration, water repellency, and soil moisture content after broadcast burning a forest site in southwest Oregon. J. Soil Water Conserv. 44, 87-90
- McNabb, D. H. and Swanson, F. J. 1990. Effects of fire on soil erosion. In: J. D. Walsad, S. R. Radosevich and D. V. Sandberg (eds.) Natural and Prescribed Fire in the Pacific Northwest Forests. Oregon State University Press, Corvallis, pp. 159-176.
- Megahan, W. F., Day, N. F., and Bliss, T. M. 1978. Landslide occurrence in the western and central northern Rocky Mountain Physiographic Province in Idaho. In: C. T. Youngberg (ed.), Forest Soils and Land Use. Proceedings of the 5th North American Forest Soils Conference. Colorado State University, Ft. Collins, pp. 116-139.
- Megahan, W. F., King, J. G., and Seyedbagheri, K. A. 1995. Hydrologic and erosional responses of a granitic watershed to helicopter logging and broadcast burning. For. Sci. 41,777-795
- Mersereau, R. C. and Dyrness, C. T. 1972. Accelerated mass wasting after logging and slash burning in western Oregon. J. Soil Water Conserv. 27, 112-114
- Meyer, G.A., and J. L. Pierce. (*this issue*). Geomorphic and climatic controls on fire-induced sediment pulses in Yellowstone and central Idaho: a holocene perspective. Submitted to For. Ecol. Manage.

Meyer, G. A., Pierce, J. L., Wood, S. H., and Jull, A. J. T. 2001. Fire, storms, and erosional events in the Idaho batholith. Hydrol. Process. 15, 3025-3038

Meyer, G. A. and Wells, S. G. 1997. Fire-related sedimentation events on alluvial fans, Yellowstone National Park, U.S.A. J. Sediment. Res. 67, 776-791

Meyer, G. A., Wells, S. G., and Jull, A. J. T. 1995. Fire and alluvial chronology in Yellowstone National Park: Climatic and intrinsic controls on Holocene geomorphic processes. Geol. Soc. Am. Bull. 107,1211-1230

Miller, D. J., C. H. Luce, and L. E. Benda. (*this volume*). Time, space, and episodicity of physical disturbance in streams. Submitted to For. Ecol. Manage.

Miller, D. W., Geraghty, J. J. and Collins, R. S. 1963. Water atlas of the United States. Water Information Center, Inc. Port Washington, L.I., N.Y.

Montgomery, D. R., Abbe, T. B., Buffington, J. M., Peterson, N. P., Schmidt, K. M., and Stock, J. D. 1996. Distribution of bedrock and alluvial channels in forested mountain drainage basins. Nature 381,587-589

Moody, J. A. 2001. Sediment transport regimes after a wildfire in steep mountainous terrain. In: Proceedings of the Seventh Federal Interagency Sedimentation Conference, Reno, Nevada. pp. X41-X48.

Moody, J. A. and Martin, D. A. 2001a. Initial hydrologic and geomorphic response following a wildfire in the Colorado Front Range. Earth Surf. Process. Landforms 26, 1049-1070

Moody, J. A. and Martin, D. A. 2001b. Post-fire, rainfall intensity-peak discharge relations for three mountainous watersheds in the western USA. Hydrol Process 15, 2981-2993

Nakamura, F., Swanson, F. J., and Wondzell, S. M. 2000. Disturbance regimes of stream and riparian systems—a disturbance-cascade perspective. Hydrol. Process. 14, 2849-2860

NOAA. undated. Rainfall Frequency Atlas of the United States. US. Dept. Commerce, National Oceanographic and Atmospheric Administration, National Weather Service, Office of Hydrology. Atlas 2, Vol. 1. (http://lwf.ncdc.noaa.gov/oa/documentlibrary/rainfall

.html) Rapp, A., Li, J., and Nyberg, R. 1991. Mudflow disasters in mountainous areas. Ambio 20, 210-218

Reeves, G. H., Benda, L. E., Burnett, K. M., Bisson, P. A., and Sedell, J. R. 1995. A disturbance-based ecosystem approach to maintaining and restoring freshwater habitats of evolutionarily significant units of anadromous salmonids in the Pacific Northwest. Am. Fish Soc. Symp. 17, 334-349

Robichaud, P. R. 2000. Fire effects on infiltration rates after prescribed fire in Northern Rocky Mountain forests, USA. J. Hydrol. 231-232, 220-229 Robichaud, P. R. and Brown, R. E. 1999. What happened after the smoke cleared: Onsite erosion rates after a wildfire in eastern Oregon. Wildland Hydrol. June/July,419-426

Robichaud, P. R. and Hungerford, R. D. 2000. Water repellency by laboratory burning of four nothrern Rocky Mountain forest soils. J. Hydrol. 231-232, 207-219

Swanson, F. J. 1981. Fire and Geomorphic Processes. In: Gen. Tech. Rep. USDA For. Serv. WO-26. Washington DC. pp. 401-420.

Swanson, F. J. and Dyrness, C. T. 1975. Impact of clearcutting and road construction on soil erosion by landslides in the western Cascade Range, Oregon. Geology 3, 393-396

Swanston, D. N. 1971. Principal soil movement processes influenced by roadbuilding, logging and fire. In: Proceedings of a Symposium: Forest Land Uses and Stream Environment. Oregon State University, Corvallis. pp. 29-40.

Swanston, D. N. 1991. Natural processes. In: Meehan, W. R., (ed.), Influences of forest and rangeland management on Salmonid fishes and their habitats. Am. Fish. Soc. Special Pub. 19. Bethesda, Maryland.

- Trimble, S. W. and Crosson, P. 2000. U.S. soil erosion rates -- Myths and reality. Science 289, 248-250
- Troendle, C. A. and C. F. Leaf. 1980. Hydrology (Chapter III). An approach to water resources evaluation of non-point silivicultural sources (A procedural handbook). U.S. Environmental Protection Agency; EPA-600/8-80-012.

Varnes, D. J. 1978. Slope movement types and processes. In: R. L. Schuster and R. J. Krizek (eds.), Landslide analysis and control. National Academy of Sciences, Transportation Research Board, Special Report 176. pp. 11-13.

Wells, C. G., Campbell, R. E., DeBano, L. F., Lewis, C. E., Fredriksen, R. L., Franklin, E. C., Froelich, R. C., and Dunn, P. H. 1979. Effects of fire on soil: A state-of-knowledge review. Gen. Tech. Rep. USDA For. Serv. WO-7. Washington D.C.

Wells, W. G. 1987. The effects of fire on the generation of debris flows in southern California. In: J. E. Costa and G. F. Wieczorek (eds.), Debris Flows/Avalanches: Process, Recognition, and Mitigation. Geological Society of America Review in Engineering Geology VII. pp. 105-114.

Wilcox, B. P., Newman, B. D., Brandes, D., Davenport, D. W., and Reid, K. 1997. Runoff from a semiarid ponderosa pine hillslope in New Mexico. Water Resourc. Res. 33,2301-2314

Wohlgemuth, P. M., Hubbert, K. R., and Robichaud, P. R. 2001. The effects of log erosion barriers on post-fire hydrologic response and sediment yield in small forested watersheds, southern California. Hydrol Process 15,3053-3066

Wondzell, S. M. and Swanson, F. J. 1999. Floods, channel change, and the hyporheic zone. Water Resourc. Res. 35,555-567